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1 **Comparative study of Cu uptake and early transcriptome responses in the green microalga**
2 ***Chlamydomonas reinhardtii* and the macrophyte *Elodea nuttallii***

3

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13

14 **Abstract**

15 Microalgae are widely used as representative primary producers in ecotoxicology, while
16 macrophytes are much less studied. Here we compared the bioavailability and cellular toxicity
17 pathways of 2h-exposure to $1 \cdot 10^{-6}$ M Cu in the macrophyte *Elodea nuttallii* and the green
18 microalga *Chlamydomonas reinhardtii*.

19 Toxicokinetics suggested that accumulation is higher in the algae than in the macrophyte, while
20 RNA-Seq revealed a similar number of regulated genes. Early-regulated genes were congruent
21 with expected adverse outcome pathways for Cu with GO terms including gene regulation,
22 energy metabolism, transport, cell processes, stress, RedOx metabolism and development.
23 However, the level of gene regulation was higher in *E. nuttallii* than in *C. reinhardtii* and several
24 categories were more represented in the macrophyte than in the microalga. Moreover, several
25 categories including OPP, nitrate metabolism and metal handling were only found for *E.*
26 *nuttallii*, while categories such as cell motility, polyamine metabolism, mitochondrial electron
27 transport and TCA were unique to *C. reinhardtii*. These differences were attributed to
28 morphological and metabolic differences and highlighted dissimilarities between a sessile and a
29 mobile species. Our data supports the efficiency of transcriptomics to assess early molecular
30 responses in biota, and the importance of studying more aquatic plants for a better understanding
31 on the impact and fate of environmental contaminants.

32
33 **Keywords:** copper; primary producers; speciation modelling; toxicokinetics; transcriptomics.

34
35 **Capsule:** Cu impact is different for microalgae and macrophytes: accumulation is higher in the
36 algae, but higher transcriptome response occurred in the macrophyte

37 **Introduction**

38 Primary producers are key organisms of aquatic ecosystems: phytoplankton sustains the largest
39 ecosystem on the Earth, contributing to about half of the primary production on our planet
40 although accounting for less than 1% of photosynthetic biomass (Bañuelos et al., 1998).
41 Macrophytes, including plants dominate primary production in shallow waters including littorals,
42 rivers, marshes, ponds and lakes (Noges et al., 2010). In addition they are key elements of the
43 aquatic ecosystems by providing support, shelter, food and oxygen to many organisms including
44 epiphytes (Thomaz and Cunha, 2010). Studying primary producers' response to a variation of the
45 concentrations of vital and toxic trace metals is thus an important step to understand and estimate
46 their impact in aquatic ecosystems. Indeed, if primary producers are affected they will also
47 indirectly influence higher trophic levels in an ecosystem through food webs (Fleeger et al.,
48 2003; Daam et al., 2009). Often microalgae are hypothesized as representative primary producers
49 based on the assumption that all organisms respond to stresses similarly (Clemens, 2006).
50 Nevertheless, seldom comparisons, for example of plants with algae (Paz et al., 2007; Beauvais-
51 Fluck et al., 2018a) or mosses (Rother et al., 2006), reveals the existence of different stress and
52 tolerance mechanisms. In such a context, the precise mechanisms of cellular handling (and
53 toxicity) are to further elucidate to better understand the similarities and differences in different
54 primary producers and anticipate and mitigate trace metals effect in the environment.

55 Copper (Cu) is an essential metal to all plants and animals. It participates in fundamental
56 physiological processes (e.g. photosynthetic electron transport, mitochondrial respiration) and is
57 a cofactor for many enzymes (e.g. superoxide dismutase, cytochrome c oxidase) (Castruita et al.,
58 2011). Due to its high reactivity, Cu concentration is tightly regulated inside cells by a complex
59 homeostasis network (Andres-Colas et al., 2006). This homeostasis network has been studied in

60 several model species and there are evidences of a high conservation throughout the evolution
61 (Burkhead et al., 2009; Page et al., 2009). However, when in excessive concentrations Cu causes
62 oxidative stress and photosynthesis inhibition due to adverse effects on the same cellular
63 processes where it is needed, such as enzyme activity and photosynthetic electron transport
64 (Monferran et al., 2009; Razinger et al., 2010; Upadhyay et al., 2011). Thus, Cu concentrations,
65 and its biological availability are important parameter for environmental quality in natural
66 environments. Elevated Cu concentrations in aquatic ecosystems are directly related to human
67 activities involving the production of industrial (e.g. pesticide use and agricultural run-off, mine
68 tailings) and domestic wastes (e.g. urbanization, automobile exhausts). Naturally occurring
69 concentrations of Cu range between 10^{-9} M to 10^{-8} M in freshwater systems, but Cu can easily
70 reach 10^{-6} M in locations receiving anthropogenic inputs such as freshwater ecosystems close to
71 vineyards or mining areas (Kupper and Andresen, 2016).

72 The present study aimed thus to compare Cu toxicokinetic and transcriptomic responses in two
73 aquatic primary producers: a macrophyte *Elodea nuttallii* and a green microalga
74 *Chlamydomonas reinhardtii*, respectively representing aquatic plants and phytoplankton
75 typically found in the benthic environment and the water column. In the present study, we
76 hypothesized that bioavailability and responses to toxic metals were similar in a microalgae and
77 a macrophyte exposed in similar experimental conditions. More in detail, we compared cellular
78 toxicity pathways of Cu in both organisms using transcriptomics (RNAseq) and determining the
79 uptake. This research will thus increase our level of understanding of the functioning of key
80 organisms, and eventually, the data produced will provide a scientific base to conduct sound risk
81 assessment of our freshwater ecosystems to preserve their high socio-economic and
82 environmental value.

83

84 **Material and methods**

85 *Labware*

86 All material was washed in 10 % HNO₃ baths, thoroughly rinsed with ultrapure water (MilliQ
87 Direct system, Merck Millipore) and dried under a laminar flow hood. Material for culture and
88 experiments, including media, were additionally autoclaved (1 bar, 121°C, 20 min) to avoid
89 microbial contamination.

90

91 *Exposure of algae and macrophytes*

92 *Chlamydomonas reinhardtii* (wild type strain CPCC11, Canadian Phycological Culture Centre)
93 were grown under axenic conditions in an incubator (Multitron Infors HT) at 20.2 ± 0.5 °C with
94 a 24 h light cycle ($110 \mu\text{mol}\cdot\text{phot}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$) and a rotary shaking (115 rpm). Cells were cultured in
95 a 4× diluted Tris-Acetate-Phosphate medium (TAP) (Rensing et al., 2008). At the mid-
96 exponential growth phase (62 h after inoculation), cells were harvested by centrifugation (10
97 min, 1300g), rinsed and re-suspended in the exposure medium at a final density of $(8.1 \pm 1.1)\cdot 10^5$
98 cells·mL⁻¹.

99 Shoots of *E. nuttallii* were collected in Lake Geneva, and a culture established and maintained in
100 microcosms as previously described (Regier et al., 2013b). Exposures were conducted on three
101 10 cm-long shoots without roots. Cultures and experiments were conducted in the laboratory
102 under the same controlled conditions (16/8 h light/dark; 1000 lux; 20 ± 1 °C).

103 Both organisms were exposed in triplicates to nominal concentration of 10^{-6} M Cu added as
104 CuSO₄ (Sigma Aldrich) in an artificial medium ($8.2\cdot 10^{-4}$ M CaCl₂, $3.6\cdot 10^{-4}$ M MgSO₄, $2.8\cdot 10^{-4}$ M
105 NaHCO₃, $1.0\cdot 10^{-4}$ M KH₂PO₄ and $5.0\cdot 10^{-6}$ M NH₄NO₃, pH 6.9 ± 0.1) during 10, 30 min, 1, 4 and

106 8 h. Organisms exposed in the absence of metal in the medium were used as control. The choice
107 of Cu exposure concentration corresponds to a sublethal concentration in similar experimental
108 conditions, e.g. resulting in the EC10 growth inhibition in *C. reinhardtii* for 24 h-long exposure
109 (Cheloni et al., 2014) and is 4× lower than the concentration resulting in a 15 % decrease of
110 chlorophyll content in *E. nuttallii* after 2h-long exposure.

111

112 *Cu uptake and modelling*

113 Cu uptake by microalgae and macrophyte was characterized by total and intracellular Cu (Cu_{int})
114 contents. Half of exposed organisms was rinsed with 10^{-3} M ethylene-diamine-tetraacetic-acid
115 (EDTA; Sigma-Aldrich, Buchs, Switzerland) prepared in the exposure medium, to rinse metal
116 surface-adsorbed or loosely bound to the cell wall and determine Cu_{int} . The other half of exposed
117 organisms was rinsed with medium without Cu (media-rinsed) and therefore represents the sum
118 of adsorbed Cu (Cu_{ads}) and Cu_{int} .

119 EDTA-rinsed and media-rinsed samples were freeze-dried (Beta 1-8 K), digested with 65%
120 HNO_3 (Suprapur Merck KGaA) at 90°C for 1 h and analyzed by inductively coupled plasma mass
121 spectrometry (ICP-MS; 7700x, Agilent Technologies). Concentration in media was measured in
122 acidified samples (0.5% v/v HNO_3 Suprapur) by ICP-MS. Cu concentration in unspiked artificial
123 medium was $2.3 \pm 0.3 \cdot 10^{-9}$ M Cu. Effective initial concentration in spiked media were $1.09 \pm$
124 $0.20 \cdot 10^{-6}$ M Cu ($1.85 \pm 0.47 \cdot 10^{-7}$ M Cu^{2+}) and $2.26 \pm 0.21 \cdot 10^{-6}$ M Cu ($3.84 \pm 0.51 \cdot 10^{-7}$ M Cu^{2+}),
125 for *E. nuttallii* and *C. reinhardtii*, respectively.

126 The Cu uptake was modelled using a first-order mass transfer model following two-compartment
127 system and equations below (eq 1 and 2):

128

129 $C_t = C_0 + \frac{a}{k(1 - e^{-kt})}$ (eq 1)

130 $a = k_1 \times C_e$ (eq 2)

131

132 where C_t is the metal concentration in cells ($\mu\text{mol}_{\text{Cu}}\cdot\text{g}^{-1}\cdot\text{dw}$) at time t (hours), k is the elimination
133 rate constant (h^{-1}) and a is the uptake flux ($\mu\text{mol}_{\text{Cu}}\cdot\text{g}^{-1}\cdot\text{dw}\cdot\text{h}^{-1}$), k_1 is the uptake rate constant
134 ($\mu\text{mol}_{\text{Cu}}\cdot\text{g}^{-1}\cdot\text{dw}\cdot\text{h}^{-1}$), C_e is the bioavailable concentration in the medium ($\mu\text{mol}_{\text{Cu}}\cdot\text{g}^{-1}\text{ dw}$), and C_0
135 is the constitutive metal concentration measured in cells at the beginning of the exposure
136 (Martins and Boaventura, 2002; Gimbert et al., 2008).

137

138 *RNA-sequencing (RNAseq) and quantification of differential gene expression*

139 Transcriptome response of *C. reinhardtii* and *E. nuttallii* exposed 2 h to Cu was assessed through
140 RNASeq (Illumina HiSeq 2500 System). Total RNA was extracted as previously described using
141 TRI Reagent (Sigma-Aldrich, Buchs, Switzerland), and libraries were prepared following
142 manufacturer's protocols (Beauvais-Fluck et al., 2016; Regier et al., 2016; Beauvais-Fluck et al.,
143 2017). For *C. reinhardtii*, reads were aligned with TopHat2 (Kim et al., 2013) to the genome
144 *Creinhardtii* 236 V.9.0 (Conesa et al., 2005). For *E. nuttallii*, reads were mapped using the
145 Burrows-Wheeler Alignment (BWA v.0.7.10) tool (Li and Durbin, 2010) on the *de novo*
146 transcriptome available for this organism (Regier et al., 2016). For both organisms reads were
147 counted using the Python package HTSeq (Anders et al., 2015). Differential gene expression
148 analysis was performed in the software CLC Main Workbench (Version 7, CLC bio, QIAGEN,
149 Denmark) based on normalized counts and EdgeR package (Robinson et al., 2010). Significant
150 differently expressed transcripts vs Control were defined with a threshold of false discovery rate
151 (FDR) <0.1%. Ontology term assignments were done using MapMan (Table S1) (Thimm et al.,

152 2004; Usadel et al., 2009). Data are available in the Gene Expression Omnibus database
153 (GSE65109).

154

155 **Results and discussion**

156 *Cu uptake by primary producers*

157 Accumulation of Cu was measured in media-rinsed and EDTA-rinsed *E. nuttallii* and *C.*
158 *reinhardtii* over time. Exposure to Cu in *C. reinhardtii* resulted in higher and faster accumulation
159 than in *E. nuttallii* (Figure 1, Table 1). Concentrations measured in media-rinsed *C. reinhardtii*
160 reached a plateau in 2h, and resulted in the highest measured bioaccumulation (up to 40 $\mu\text{mol}\cdot\text{g}^{-1}$
161 dw). The uptake of *C. reinhardtii* was of the same order of magnitude than reported in previous
162 studies with *C. reinhardtii* and other green freshwater algae (Stoiber et al., 2012). For
163 comparison, media-rinsed *E. nuttallii* showed a bioaccumulation of close to 8 $\mu\text{mol}\cdot\text{g}^{-1}$ dw at 8h,
164 and no obvious evidence of a plateau. Moreover, after 2h exposure *E. nuttallii* internalized
165 (EDTA-washed) $1.33 \pm 0.31 \mu\text{mol}\cdot\text{g}^{-1}$ dw ($1.04 \pm 0.24 \mu\text{mol}\cdot\text{g}^{-1}$ dw in control), while *C.*
166 *reinhardtii* internalized $4.69 \pm 0.18 \mu\text{mol}\cdot\text{g}^{-1}$ dw ($1.44 \pm 0.06 \mu\text{mol}\cdot\text{g}^{-1}$ dw in control). Data
167 suggested that internalization is higher and faster in the algae than in the macrophyte. This
168 difference can be attributed to the fact that the full surface of the unicellular algae is in contact
169 with the media, whereas in the macrophyte only the external layer of cells is directly exposed,
170 most certainly resulting in a gradient of metal concentrations between cells. Besides, the surface-
171 to-volume ratio is much higher in an unicellular organism and thus is expected to result in higher
172 uptake (Lindemann et al., 2016). However, proportion of Cu accumulated in cell walls was
173 higher in *C. reinhardtii* than in *E. nuttallii*, suggesting that adsorption of Cu was predominant in
174 *C. reinhardtii* and/ or EDTA-washing procedure was more efficient. Cell walls are known to

175 play a central role in plant and microalgal tolerance to metals: for example, 50 % of Cu was
176 accumulated in the cell walls in *Cystoseira tamariscifolia* (Celis-Pla et al., 2018), 20% was
177 adsorbed (or EDTA-extractable) for *Chlorella kessleri* (Lamelas et al., 2009). In the charophyte,
178 *Nitellopsis obtusa* exposed 3h to both Cu-nanoparticles or CuSO₄, the major part of Cu
179 accumulated in cell walls (Manusadzianas et al., 2017). Similarly, a previous study in *E. nuttallii*
180 measured an increased proportion over time of cadmium and mercury in cell walls,
181 concomitantly with an increased lignification of cell walls after 7 d exposure (Larras et al.,
182 2013).

183 The one-compartment model well fitted Cu accumulation in *E. nuttallii* and *C. reinhardtii*. In
184 both organisms significant and similar (*a*) and (*k*) were estimated by the model normalized by
185 the effective concentration in media. Modelling further allowed estimating that the steady state
186 was approached in less than 2h for *C. reinhardtii*. Similar, fast uptake and plateau has been
187 observed in *C. reinhardtii* exposed to increasing concentrations of Cu⁶⁵ (Jamers et al., 2013).

188

189 *Transcriptomic response*

190 In total 1397 and 1258 genes were regulated by 2h exposure to 1·10⁻⁶ M Cu in *C. reinhardtii* and
191 *E. nuttallii* respectively. The similar number of regulated genes, used as a proxy of stress,
192 suggested that both *E. nuttalli* and *C. reinhardtii* faced a similar level of stress (Dranguet et al.,
193 2017). However, among those, 841 (67%) and 624 (44%) genes were upregulated, while 417 and
194 773 were down-regulated in *E. nuttalli* and *C. reinhardtii*, respectively. Besides, the level of
195 gene regulation was higher in *E. nuttallii* (log₂FCrange= 17.9) than in *C. reinhardtii*
196 (log₂FCrange= 8.8), suggesting a higher impact of Cu in the macrophyte than the microalgae, in

197 line with similar observations made for Hg in controlled and in the field exposure comparing the
198 same species (Dranguet et al., 2017; Beauvais-Fluck et al., 2018a).

199 In term of abundance of GO terms for both species a predominant part of regulated genes had
200 unknown function (62% for *C. reinhardtii* and 33 % for *E. nuttalli*; Figure 3) indicating
201 considerable potential for new discovery in the biology of Cu. Enriched pathway analysis
202 revealed that Cu exposure regulated genes involved in gene regulation (i.e. RNA, protein,
203 signaling) and energy metabolism (Figure 3). Other categories included genes involved in
204 transport, cell processes, hormone metabolism, stress, RedOx metabolism and development
205 (Figure 3). Overall, results support a response of both species to avoid stress (e.g. oxidative
206 stress) and effects on development/growth and nutrition with a significant modification of the
207 energy metabolism. Regulated genes were thus in line with expected adverse outcome pathways
208 for Cu, i.e. impact on photosynthesis, RedOx, growth and nutrition, although only 2h exposure
209 was performed. This confirms the potential of transcriptomics to reveal early-responses at
210 environmental concentrations (Regier et al., 2013a; Dranguet et al., 2017; Beauvais-Fluck et al.,
211 2018b). Not surprisingly this short exposure resulted in few physiological endpoints
212 significantly different vs control (Table S2 and S3) (Jamers et al., 2013; Jiang et al., 2016). More
213 in detail, here photosynthesis efficiency is reduced in *C. reinhardtii* by 7% (Table S3) and class
214 III peroxidase activity (POD) is reduced 50× in *E. nuttalli* (Table S2). In the same line, a
215 previous study in *C. reinhardtii* revealed that exposure to a similar free ion concentration 10^{-7} M
216 Cu^{2+} , induced Glutathione Peroxidase genes after 2h and reduced growth after 24h, although no
217 cellular impact was measured including membrane permeability, reactive oxygen species
218 production and lipid peroxidation (Cheloni et al., 2014).

219 A previous study showed that exposure of *E. nuttallii* to 10^{-6} M Cu reduced superoxide
220 dismutases activity after 1h and reduced root growth after 24h, but had no significant effect on
221 chlorophyll content, photosynthesis efficiency and class III peroxidase activity (Regier et al.,
222 2015). Similar observation has been obtained with Cu toxicity in *C. reinhardtii*: exposure to
223 excess Cu induced ROS production and antioxidative response in *C. reinhardtii* (Jamers et al.,
224 2006; Stoiber et al., 2013; Jiang et al., 2016). In the same line, a study on the rootless submerged
225 shoots of *Ceratophyllum demersum* exposed 6 weeks to a range of concentrations between 10^{-9} -
226 10^{-7} M Cu, showed that nutrient uptake/distribution, photosynthesis efficiency and chlorophyll
227 content were affected by Cu (Thomas et al., 2013). Nutrition is impacted because an excess Cu
228 competes with the various essential metals according to the Irving–William series and induces
229 deficiency of essential ions (Mg^{2+} , Zn^{2+} , etc.) (Mosulen et al., 2003) and impairment of
230 metalloprotein functioning. However, although all toxic metals might induce the same core stress
231 related changes on genes, transcriptome analysis has resulted in the identification of genes
232 specific to each metal (Kovalchuk et al., 2005; Weber et al., 2006; Simon et al., 2008).
233 Nonetheless, data have rarely been compared between species exposed in similar experimental
234 settings (Dranguet et al., 2017; Beauvais-Fluck et al., 2018a).

235 Here, several categories were more represented in the macrophyte than in the microalga (Figure
236 2), including stress (abiotic), development, cell vesicle transport, hormone metabolism (abscisic
237 acid, ethylene, jasmonate), cell wall (cellulose, hemicellulose and pectin synthesis), secondary
238 metabolism (phenylpropanoid, wax, flavonoids), and transport (transport P- and V-ATPases,
239 Major Intrinsic Proteins, nitrate). The present data for *E. nuttallii* were in agreement with a
240 microarray analysis in roots of rice exposed 3h to $5 \cdot 10^{-6}$ M Cu, notably concerning dysregulation
241 of genes involved in vesicle transport, flavonoids metabolism and jasmonate (Lin et al., 2013).

242 Authors further showed by knockout of genes necessary for this vesicle transport and exposure
243 of roots to vesicle trafficking inhibitors, that Cu interacts with vesicle transport and that this
244 vesicle transport is essential for signaling via ROS for activating defenses (Lin et al., 2013).
245 Results of the present study allowed to propose a possible model of cellular mechanisms
246 involved in Cu detoxification and protection in *E. nuttallii*: Cu increases intracellular transport,
247 e.g. vesicle trafficking and ABC transport, and induces a flavonoid-mediated detoxification
248 pathway. In addition, the toxicity mechanisms such as JA biosynthesis and cellular component
249 biogenesis were regulated in response to Cu exposure. In comparison, the categories of OPP,
250 nitrate metabolism and metal handling were absent in Cu regulated genes in *C. reinhardtii*.
251 Conversely, cell motility, DNA, polyamine metabolism, mitochondrial electron transport, and
252 TCA categories were found in *C. reinhardtii*, while absent in *E. nuttallii*. Moreover, the level of
253 regulation of the categories found in common in both species was higher in *E. nuttallii* than in *C.*
254 *reinhardtii*. These differences certainly highlight the dissimilarities between basal and
255 background metabolism in two different species, as well as between a sessile and a mobile
256 organism (Dranguet et al., 2017). Besides, genome sequencing has revealed that *C. reinhardtii*
257 possesses numerous genes derived from the last plant-animal common ancestor that have been
258 lost in angiosperms, including transporters and the possibility of extensive metabolic flexibility
259 (Merchant et al., 2007). Taken together, our divergent observations on how an unicellular and a
260 multicellular organism take up and are impacted by Cu may imply that homeostasis networks are
261 more species-specific than generally thought.

262 We further found several differences at the level of subcategories. For example, in the
263 ‘Photosynthesis’ category, genes of *C. reinhardtii* were mainly involved in the light reaction, in
264 particular photosystem I (PSI), while in *E. nuttallii* genes were involved both in PS I and PS II,

265 as well as photorespiration, suggesting that the photosynthesis was impacted more widely by Cu
266 toxicity in the macrophyte. Generally, Cu has been reported to impact more PS II than PS I in
267 plants. In PS II, the reaction center and LHC II by substitution of Mg²⁺ in its chlorophyll have
268 been shown to be targets of Cu toxicity (Kupper et al., 1996; Kupper and Andresen, 2016). In the
269 macrophyte *C. demersum* nanomolar concentrations of Cu affected the PS II reaction center
270 (Thomas et al., 2013). In this regard, our finding of Cu impact on PSI in *C. reinhardtii* is striking
271 and might point to structural differences between photosystems as well as background defense
272 pools in the studied species (Castruita et al., 2011).

273

274 **Conclusion**

275 Overall, the exposure to 10⁻⁶ M Cu resulted in different cellular toxicity pathways in a microalga
276 and a macrophyte. This fact together with the distinct exposure routes of the benthic macrophyte
277 and lentic microalgae suggest that similar Cu concentrations might affect differently both species
278 in the ecosystem. Defining ecological thresholds of adverse outcomes for environmental
279 contaminants represents a critical component of chemical assessment and management
280 programs. Our data call for including more species of aquatic plants for determining ecological
281 thresholds for environmental contaminants. In this context, transcriptomics is confirmed as a
282 useful tool to assess early responses at environmental concentrations.

283

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291

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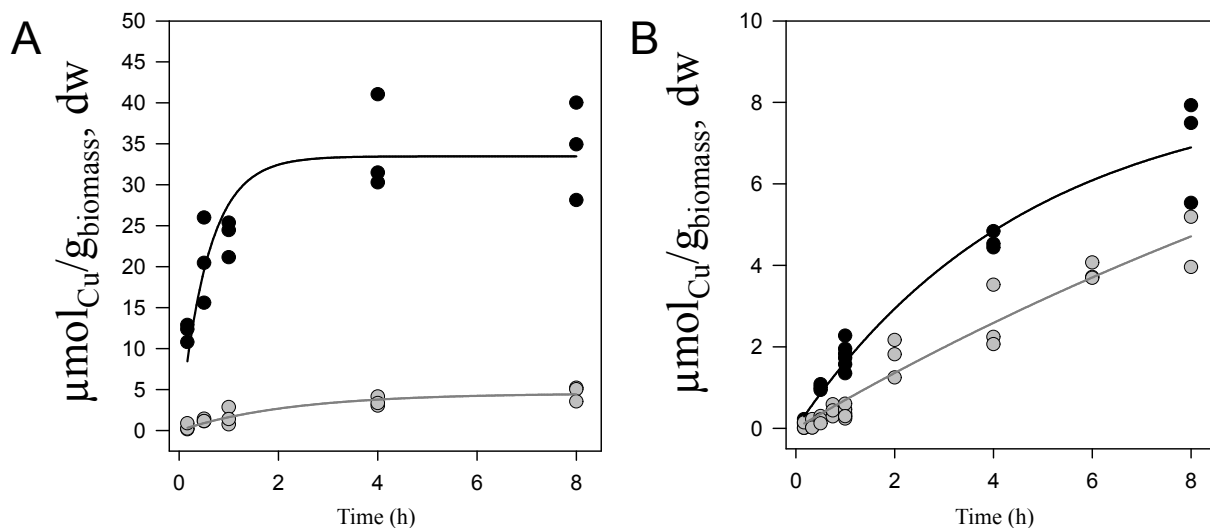
477 **Table 1:** Modelled parameters of Cu uptake in EDTA-rinsed *E. nuttallii* or *C. reinhardtii*
478 exposed to $1 \cdot 10^{-6}$ M Cu. Uptake flux (a) and elimination rate constant (k) of Cu were divided by
479 the effective concentration of metal at beginning of the test to allow inter species comparison and
480 are thus presented as a' and k' .

	<i>E. nuttallii</i>	<i>C. reinhardtii</i>
a' (L·g ⁻¹ ·h ⁻¹)	0.74 ± 0.87	0.32 ± 0.05
k' (μmol·L ⁻¹ ·h ⁻¹)	0.06 ± 0.08	0.07 ± 0.04
R ²	0.93	0.85

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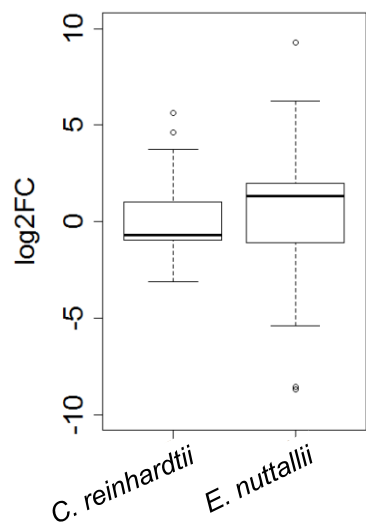
485 **Figure 1:** Cu toxicokinetics in *C. reinhardtii* (A) and *E. nuttallii* (B) exposed to $1 \cdot 10^{-6}$ M Cu.

486 Organisms were EDTA-rinsed (grey) or media-rinsed (black) to differentiate between adsorbed
487 and internalized metal.

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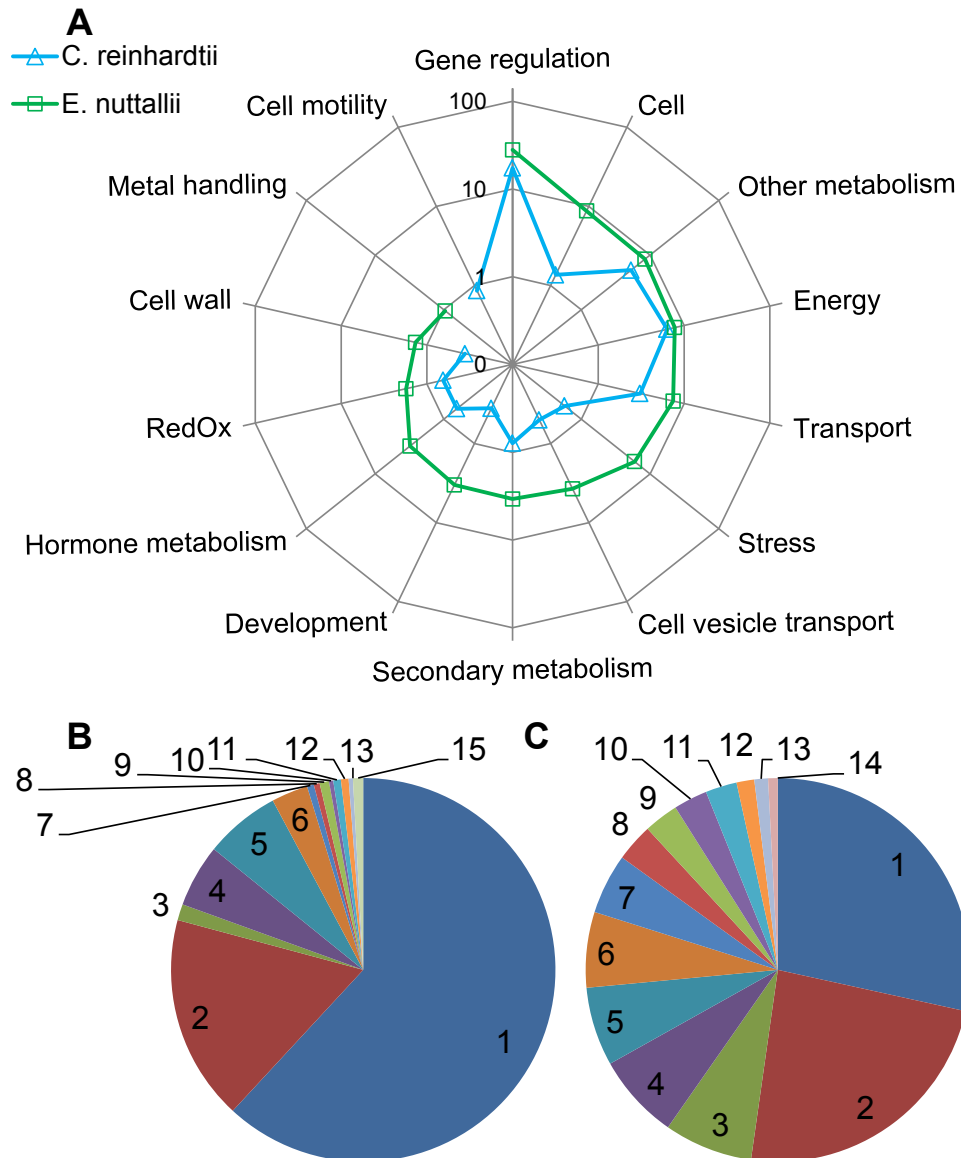
491

492 **Figure 2:** Fold-changes (log2FC) of significant regulated genes in *C. reinhardtii* and *E. nuttallii*

493 exposed 2h to $1 \cdot 10^{-6}$ M Cu.

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496 **Figure 3:** Number of significant dysregulated genes (DG) annotated in the respective enriched
 497 biological pathways (MapMan) in *C. reinhardtii* (triangles) and *E. nuttallii* (square) exposed 2h
 498 to Cu (A). Proportion (%) of main functional GO categories (MapMan) of dysregulated genes
 499 in *C. reinhardtii* (B) and *E. nuttallii* (C) exposed 2 h to Cu (1: Unknown, 2: Gene regulation, 3:
 500 Cell process, 4: Other metabolism, 5: Energy metabolism, 6: Transport, 7: Stress, 8: Cell vesicle
 501 transport, 9: Secondary metabolism, 10: Development, 11: Hormone metabolism, 12: RedOx, 13:
 502 Cell wall, 14: Metal handling, 15: Cell Motility).

SUPPORTING INFORMATION

Comparative study of Cu uptake and early transcriptome responses in the green microalga

Chlamydomonas reinhardtii* and the macrophyte *Elodea nuttallii

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Table S1: MapMan ontology enrichment analysis for significantly dysregulated transcripts measured in *Chlamydomonas reinhardtii* and *Elodea nuttallii* exposed 2 h to Cu vs control (XLS file in attachment).

Table S2: Effect on pigments content (chlorophyll; anthocyanin) and oxidative stress enzymes activity (POD, class III peroxidase; SOD, superoxide dismutase) in *E. nuttallii* exposed 2 h to Cu or not (control) (results are presented as % of control; bold characters indicate significant difference with control, n = 3 ± SD; T-test *p*-value < 5%; fw, fresh weight).

<i>E. nuttallii</i>	Chlorophyll (mg/g fw)	Anthocyanin (µg/g fw)	POD activity (nkat/mg _{protein})	SOD activity (U/µg _{protein})
Control	100.0 ± 15.2	100.0 ± 15.1	100 ± 20.1	100.0 ± 7.9
Cu	87.0 ± 16.5	110.7 ± 7.1	2.4 ± 0.3	108.9 ± 5.9

Table S3: Effect on chlorophyll content and efficiency, and oxidative stress biomarkers (membrane integrity, ROS) in *C. reinhardtii* exposed 2 hours to Cu or not (control) (results are presented as % of control; bold characters indicate significant difference with control, n = 3 ± SD; T-test *p*-value < 5%).

<i>C. reinhardtii</i>	chlorophyll a	Fv/Fm	membrane integrity (% cell affected)	ROS generation (% cell stained)
Control	100 ± 6.5	100 ± 1.7	10 ± 1.5	10.0 ± 1.5
Cu	91.9 ± 3.3	93.2 ± 1.0	12.5 ± 1.0	12.5 ± 1.0